

1 **Short rotation coppice of leguminous tree *Leucaena* spp. improves soil fertility while producing**
2 **high biomass yields in Mediterranean environment**

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4 M. Fernández^{1*}, J. Alaejos¹, E. Andivia^{1,2}, P. Madejón³, M.J. Díaz¹, R. Tapias¹

5 ¹ Escuela Técnica Superior de Ingeniería. Universidad de Huelva. Avda. Fuerzas Armadas s/n, 21071,
6 Huelva (Spain).

7 ² Present address: Departamento de Biodiversidad, Ecología y Evolución. Universidad Complutense
8 de Madrid. C/ José Antonio Novais, 12, 28040, Madrid (Spain)

9 ³ Instituto de Recursos Naturales y Agrobiología de Sevilla (IRNAS), CSIC, Avenida Reina Mercedes
10 10, P.O. Box 1052, 41080, Sevilla (Spain).

11 * Corresponding author: manuel.fernandez@dcaf.uhu.es ; Tf. 34 959217712; Fax 34 959217560

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13 **Abstract:** The use of woody nitrogen-fixing plant species as multipurpose and energy crops aims to
14 enhance biomass yield while improving soil properties. Yet, the effectiveness of this option is still
15 under debate especially the use of short rotation cropping in water and nutrient limited environments.
16 This study investigated whether short rotation coppicing of four taxa of multipurpose biomass woody
17 legume *Leucaena* spp. can improve soil conditions when grown for biomass under a Mediterranean
18 environment. Biomass yield, mineral composition and heat value of the biomass and the mineral and
19 organic matter content of soil were evaluated. Under favorable growing conditions, woody dry
20 biomass production was up to 29 Mg ha⁻¹ year⁻¹ with slight but significant differences between the four
21 taxa that were tested. After 11 years of cropping, the soil showed higher fertility and microbial activity
22 compared to the uncropped plot. *Leucaena* cultivation increased soil nitrogen by 35%, dehydrogenase
23 activity by 98%, and organic matter and carbon content (by 41%). Annual cuttings resulted in the
24 highest biomass production, followed by two and three year cuttings. The mineral composition and the
25 calorific capacity of woody biomass make it suitable for commercial use as an energy source
26 (generating on average 151 MWh ha⁻¹ year⁻¹). In conclusion, short rotation coppicing of nitrogen-
27 fixing woody species results in high biomass production rates with the restoration of degraded soils,
28 constituting a sustainable agroforestry system for rural areas.

29

30 **Keywords:** nitrogen-fixing trees, energy use, pellets, litterfall, mineral nutrients.

31

32 **1. Introduction**

33 International commitments aimed at mitigating climate change lead to the promotion of clean
34 and renewable energy sources, including biomass (United Nations, 2016). Worldwide energy
35 production from biomass currently accounts for more than 60% of the renewable sources, with
36 renewable sources contributing up to 19.3% of the global energy output (REN21, 2017). In this
37 context, the use of lignocellulosic biomass, is steadily expanding. However, as the traditional
38 agroforestry systems can not supply the growing demand, plantations of fast-growing woody species
39 are necessary (Christersson and Verma, 2006; Bogdanski et al., 2010; IEA, 2016). Increasing the
40 productivity of biomass crops would reduce the amount of land needed for biomass production and
41 would increase the economic benefits of rural economies.

42 Lignocellulosic energy crops are expected to cover from 5 to 10% of the global forest area by
43 2050, mostly destined to marginal areas with degraded low fertility soils, and they will be required to
44 provide a large amount of biomass in a cost-effective manner. Thus, it is necessary to evaluate the
45 environmental impact of this intensive production system on already poor and degraded soils (Parrotta,
46 1999; Isaac et al. 2003; Vanguelova and Pitman, 2011; Lauri et al., 2014).

47 Currently, different plant species are used globally as short rotation energy crops (such as the
48 genera *Populus*, *Salix*, *Eucalyptus*, *Miscanthus*, *Cynara*, *Robinia*, *Casuarina* and *Leucaena*) (Parrotta,
49 1999; Halford and Karp, 2011; Vanbeveren et al., 2017), some of which are considered multiple-
50 purpose species providing numerous goods and services (e.g. forage, energy, soil restoration, pulp for
51 the paper industry, and construction). In this regard, the incorporation of nitrogen-fixing plant species
52 into energy crops, such as *Leucaena* spp., may represent an interesting option and, apart from the
53 traditional multiple use of this genus, the extraction of new products by biorefinery techniques
54 increases the possibility of biomass valorization (Feria et al., 2011; Loaiza et al., 2017). Nitrogen-
55 fixing species are expected to improve soil fertility while enhancing crop productivity (Parrotta, 1999;
56 Tadros et al., 2012; Conrad et al., 2017) and providing other ecosystem services, such as green fodder,
57 soil restoration and carbon sequestration (Srinivasarao et al., 2014; Mukhopadhyay et al., 2016; Gusha
58 et al., 2017). Moreover, they can contribute to decreased use of nitrogen (N) fertilizers, consequently

59 reducing energy costs (Audsley, 2003; Giuntoli et al, 2017). *Leucaena leucocephala* and *Leucaena*
60 *diversifolia*, for instance, are tropical woody legume species which are fast growing and well adapted
61 to a wide range of climate and soil types, but mainly under non-arid subtropical climate (Bray and
62 Sorensson, 1992; Parrotta, 1992, Zayed et al., 2018). *Leucaena leucocephala* is the *Leucaena* sp. most
63 extensively cultivated worldwide, spreading throughout Latin America, Asia, Africa and Oceania, but,
64 as far as we know, *Leucaena* sp. has not been widely used in commercial plantations under a
65 Mediterranean climate. However, the global distribution of plantations and research plots also includes
66 subtropical climate areas in Australia, Chile, South Africa and Spain (Zayed et al., 2018), many of
67 them under a typical Mediterranean subtropical temperate climate (i.e. *Csa* climate in the Köppen-
68 Geiger climate classification system) such as areas near the coast of the Iberian Peninsula, Italian
69 Peninsula, Greece, Turkey and North Africa (Beck et al. 2018). In Europe alone, the degradation of
70 farmland is such that 45% contains low organic matter content (< 2%), 15% presents problems due to
71 excess of inorganic nitrogen fertilization, and more than 147 million hectares suffer from serious
72 erosion problems (FAO, 2015). At least 20% of all these degraded lands are under a climate where
73 *Leucaena* spp. can grow. As the *Leucaena* genus includes fast growing multiple-purpose species, the
74 main impacts that could be expected on locations where *Leucaena* spp. are planted or have the
75 potential to be planted for biomass production are: soil fertility recovery by improving organic matter
76 and nitrogen content (Parrotta, 1999; Peng et al., 2013; Conrad et al., 2018), biofuel production for
77 heat energy (Chaplot, 2014; Loaiza et al., 2017; Dalzell, 2019), soil erosion control and therefore
78 reduced run-off and improved water quality (Shelton and Dalzell, 2007; Adhikary et al., 2017), and
79 high quality fodder for ruminant livestock (Gusha et al., 2017; Leketa et al., 2019; Santana et al.
80 2019). All this can reactivate the rural economy while providing environmental benefits. In addition,
81 cultivated *Leucaena* spp. can provide raw materials for wood and paper industries and for biorefineries
82 (Malik et al., 2004; Ferial et al., 2011; Sa'ad et al., 2019; Zayed et al., 2019), as well as restore
83 contaminated soils (Mukhopadhyay et al., 2016). Finally, the ability of *Leucaena* spp. of fixing N and
84 producing biomass for energy use contribute to reduce nitrogen fertilization and green-house gas
85 emissions, and to improve carbon sequestration (Srinivasarao et al., 2014; Conrad et al., 2017, 2018).
86 The latter, would be useful for European Union in order to comply with the promotion of the use of

87 renewable energy (European Union, 2018), and with the 2015 Paris Agreement on Climate Change
88 (United Nations Climate Change Conference, COP21).

89 Commercial *Leucaena* sp. plantations yield from 1 to 50 Mg ha⁻¹ year⁻¹ of dry biomass,
90 depending on the growing conditions. Water stress, winter frosts and soil fertility have been identified
91 as the main limiting factors for the growth and survival of these plantations (Parrotta, 1999; Tadros et
92 al., 2012; Chaplot, 2014). However, as *Leucaena* spp. can become an invasive alien species in certain
93 areas, it must be managed with caution (Chiou et al., 2016; Mello and Oliveira, 2016).

94 This study aims to evaluate whether short rotation coppicing of *Leucaena* spp. can produce a
95 large amount of high quality biomass at the same time as improving soil conditions. It has been carried
96 out in a Mediterranean environment. Soil properties were assessed under four commercial taxa of
97 *Leucaena leucocephala* (Lam.) De Wit and *Leucaena diversifolia* (Schlecht.) Benth. subjected to
98 different biomass harvest frequencies. Different taxa were used because *Leucaena* species and
99 varieties vary considerably in terms of tolerance to environmental stress and nutrient requirements
100 (Foroughbakhch et al., 2007; Sangram and Keerthika, 2013; Katunga et al., 2014). The objectives of
101 this study are: *i*) to assess plant growth and productivity of *Leucaena* spp. subjected to different short
102 rotation coppice treatments in a Mediterranean environment; *ii*) to rate their suitability for use as
103 energy crops; and *iii*) to determine the evolution of soil conditions after a cropping period of 11 years.

104

105 **2. Materials and Methods**

106 **2.1. Plant material and experimental design**

107 Seeds of two *Leucaena* species were used (two provenances – ‘Honduras’, ‘India’ – and one
108 variety – ‘K360’ – of *L. leucocephala*; and one provenance of *L. diversifolia* – ‘Hawaii’ –), all of them
109 being high productivity commercial taxa. At the beginning of March 2002, the seeds were soaked and
110 inoculated with bacteria of the genus *Rhizobium*. The following day, the seeds were sown in 300-cm³
111 containers filled with *Spagnum* peat:vermiculite (3:1 volume ratio). After three months in a nursery, in
112 the first week of June, the seedlings were planted in the field. The site was located in Huelva (Spain),
113 which has a Mediterranean climate with mild winters and a summer drought period of 3–4 months
114 (UTM zone 29S, X: 684824, Y: 4119171, 18 m a.s.l.). Mean temperature and annual rainfall in the

115 area (based on data for the previous 20 years) were 16.2 °C and 564 mm, respectively. The
116 experimental design comprised a plot, divided in two parts, which were separated 150 m from each
117 other. Plant growth and survival were evaluated in the two parts of the plot, while chemical parameters
118 of the soil and plants were measured in one of them. Each one of the four taxa was replicated four
119 times in each part. Each replicate consisted of two parallel rows, each containing eight plants, with a
120 spacing of 0.6 m between plants and 1.8 m between lines (17.28 m²); the crop density was 9259 plants
121 ha⁻¹ (Figure 1). Thus, the experimental design resulted in a total of 512 plants (4 taxa × 4 replicates ×
122 16 plants × 2 parts). In the first part of the plot, two adjacent uncropped areas with natural vegetation
123 (predominantly composed of clear scrubland containing species belonging to the genus *Cistus*) were
124 selected as controls for soil sampling at the end of the assay (Figure 1).

125 [Figure 1 here]

126 The two plot parts showed similar physical-chemical soil properties. Soils were permeable
127 with a sandy loam texture and were free of active limestone. At planting, the physical-chemical
128 properties of the surface horizon (0–20 cm depth) of the first part of the plot were as follows: 76.5% as
129 fine soil (< 2 mm particle-size) composed of 7.0% as clay, 7.5 % as silt and 85.5 % as sand, 1.1%
130 organic matter (OM), pH = 6.0 (water pH; i.e. pH_{H2O}), EC = 3.94 mS m⁻¹, 360 mg kg⁻¹ of N, 14.7 mg
131 kg⁻¹ of available P, 120 mg kg⁻¹ of available K, and bulk density of 1.5 kg/dm³. Soil preparation for
132 planting consisted of the eradication of pre-existing vegetation, followed by a deep ripping at 50 cm
133 and a shallow tillage. Since the viability of the *Leucaena*-nitrifying bacteria symbiosis was not assured
134 because they were new species in this soil, and in order to facilitate plant growth during the first
135 months after planting in this nutrient-poor soil, it was decided to apply a standard fertilization before
136 planting with N (160 kg ha⁻¹), K (110 kg ha⁻¹), P (19 kg ha⁻¹) and Mg (12 kg ha⁻¹), equivalent to a
137 11–3–9 ratio fertilizer (N–P₂O₅–K₂O), plus 1,4 MgO. Thereafter, no fertilizer was added, and eight
138 months after planting root nodules were observed in the field plot. Weeds were manually controlled
139 during the first year. From the second year onwards, weed control was unnecessary because of the
140 high cultivation density and the rapid growth of the new sprouts. Irrigation was applied only during
141 the first three months after planting to ensure plant survival. Irrigation amounted to 54 mm (4.5 mm

142 per week) from late June to mid-September. The adjacent unmanaged areas did not receive any
143 agronomic inputs (irrigation, fertilization, weeds control, etc.), they were left to their natural
144 development. Climatic conditions during the studied period are shown in Table 1.

145 [Table 1, here]

146

147 **2.2. Shoot growth and biomass production**

148 Height (H) and stem diameter (D , measured 7 cm above ground) were periodically measured
149 in 12 plants per replicate. During the first three years of the study period, the effect of the short
150 rotation cutting cycles was analyzed by measuring plant growth at one, two and three years after
151 planting. We also measured growth in re-sprouted shoots one year after cutting. For this, the first year
152 after planting plants in two replicates were cut, while plants in each of the remaining two replicates
153 were cut the second and the third year, respectively (Figure 2). Therefore, during the first year there
154 was no difference between cutting treatments, but they differed from the second year onwards. Each
155 treatment is identified according to the following convention: $R_{i,j}S_k$ indicates age in years of the
156 stump-roots (i), the number of harvests a stump has experienced (j) and the age of the aboveground
157 plant part at the time of cutting (k). For example, for the first three years studied, the *111* treatment
158 was composed by $R_{11}S_1 + R_{22}S_1 + R_{33}S_1$; the *21* treatment by $R_{21}R_2 + R_{32}S_1$, and the *3* treatment by
159 $R_{31}S_3$.

160 Additionally, on a monthly basis, phenology was evaluated for each taxon and cutting
161 treatment by recording the status of the sprouts (developing or resting) and the presence of flowers and
162 fruits (developing or ripe). Furthermore, during the summer of the second and third cropping seasons
163 (Figure 2), leaf area index (LAI) was measured using a portable radiometer (LP-80, Decagon Devices,
164 Inc., WA, USA), predawn leaf water potential (Ψ) was determined using a Scholander pressure
165 chamber (PMS Inc., Corvallis, OR, USA) and specific leaf area (SLA) was recorded ($m^2 kg^{-1}$).

166 [Figure 2 here]

167 After the third year, cultivation was extended for eight more years to determine the shoot
168 sprouting vigor by measuring TDW production, and evaluate the effect of long-term cultivation on soil

169 conditions. During that time, two replicates per plot were harvested every year and the other two were
170 harvested every two years.

171 Biomass assessment was carried out during the winter season (February to March) by cutting
172 and weighing shoots. Allometric equations were developed for those shoots that were not cut. These
173 allometric equations related to the diameter of the base (D) and/or the height of the main stem (H) with
174 the aboveground dry weight (Noulekoun et al., 2018). For this, stems with D ranging from 11 to 125
175 mm were used. Woody (W) and non-woody (NW) parts were oven-dried at 80 °C and weighed to
176 estimate dry weight (WDW and NWDW). Non-woody biomass was composed of leaves plus
177 reproductive structures and twigs less than 5.0 mm in diameter. Total dry weight (TDW) was
178 calculated as the sum of WDW and NWDW. A total of 16 different equations were tested (including
179 linear, power, exponential, polynomial, logarithmic and quadratic) with D , H , and D^2 or combinations
180 of these as predictor variables. Most of the equations yielded excellent results in terms of predicting
181 aboveground dry weight. The power equation using D as a predictor showed a very good fit ($TDW = a$
182 D^b) and was thus used for estimating the weight of aboveground biomass ($TDW = 3.803 D^{1.664}$, $R^2 =$
183 0.851 , $p < 0.001$, $n = 48$ for *L. diversifolia*; $TDW = 3.357 D^{1.614}$, $R^2 = 0.913$, $p < 0.001$, $n = 68$ for *L.*
184 *leucocephala*; TDW is expressed in g and D in mm).

185

186 **2.3. Physical-chemical characterization of plant and soil material**

187 In addition to the initial soil sampling, soil from part one of the plot was also sampled in
188 summer 11 years after planting. Two soil samples were taken between planted lines of each replicate
189 using a soil core (2 samples \times 4 replicates \times 4 taxa = 32 soil samples). Samples were then separated
190 into organic layer (i.e., litterfall) and two soil mineral layers (0–20 cm and 20–40 cm depth). Soil
191 samples for each replica and layer were then pooled. We also sampled soils in the adjacent uncropped
192 area of the plot (under natural vegetation) following the above-mentioned sampling protocol. The litter
193 layer of the soil (O horizon) showed two visually different strata; an upper deposit composed of
194 leaves, shoots and reproductive organs that were scarcely decomposed, and a lower stratum consisting

195 of plant tissues which were partially or completely decomposed. The two strata were separated for
196 thickness and dry weight quantification.

197 Litterfall was washed, oven-dried at 80 °C for two days, ground and stored in sealed containers
198 at room temperature (15–20 °C). Soil samples were divided in two subsamples. One was used for
199 physical-chemical analysis. For this, the subsample was dried at room temperature during two weeks
200 and then milled and sieved (2-mm mesh). The other subsample was used for biochemical analysis, so
201 the moisture content at the time of sampling was preserved. For this, subsamples were sieved (< 2
202 mm) and kept in the dark at a low temperature (4 °C). Nutrient content of litterfall as well as physical-
203 chemical and biochemical properties of soil were determined using standardized methods (Table 2).

204 Regarding the aboveground plant biomass, four shoots per replica were sampled, separated
205 into leaves and woody stems, and samples from each replica then pooled for analysis. They were also
206 washed, oven-dried at 80 °C, ground and stored in sealed containers at room temperature (15–20 °C),
207 and their physical-chemical properties analyzed (Table 2). Woody material was ground (Woodstock
208 3PH, Smartec®, Italy) and sieved to a particle size of 0.2 to 5.0 mm in order to create homogenous
209 samples. Then, wood pellets were manufactured using a pelleting press (PLT-400, Smartec®, Italy).
210 The sample (i.e., sawdust) moisture content was set to 128 g kg⁻¹ with a bulk density of 258 kg m⁻³ and
211 an operating temperature ranging between 95 and 105 °C. The die channels had a diameter of 6 mm;
212 the first part had a cone-shaped opening 3.5 mm deep and 70° angles; the active part was 28 mm long,
213 so the compression ratio was 4.67. Afterwards, the mechanical durability, moisture content, length and
214 diameter, and the bulk density of the pellets were determined according to methods described in ISO
215 17225-2:2014.

216 [Table 2 here]

217

218 **2.4. Data analysis**

219 Linear mixed models were used to evaluate plant growth (*D* and *H*) and dry weight (TDW,
220 WDW and NWDW) during the first three years after planting. The year (*Y*), cutting treatment (*CT*)
221 and taxa (*T*) were considered as fixed effects in the model. The triple and the pairwise interactions

222 between main effects were also included as fixed effects in the model. We considered replicate unit as
223 random term in the model to account for non-independency among observations within the same
224 individual (i.e. repeated measurements) (Zuur et al. 2009). An autoregressive correlation structure to
225 remove the first-order autocorrelation between observations was also used (Pinheiro et al. 2018).
226 Analogously, we used the same model approach to evaluate TDW for the period between the fourth
227 and the eleventh year after planting. Since we only considered two cutting treatments (annual and
228 biennial), we evaluated TDW in the fifth, seventh, ninth and eleventh years after planting. To do this,
229 we calculated TDW as $\text{Mg ha}^{-1} \text{ year}^{-1}$ by averaging biennial and annual production for the evaluated
230 period (i.e. every two years). The percentages of variance explained by fixed and random effects were
231 evaluated following Nakagawa and Schielzeth (2013). Soil parameters were analyzed by a two-way
232 ANOVA but considering cultivation treatment (*CU*: natural vegetation, *L. diversifolia*, and each of the
233 three taxa of *L. leucocephala*) and soil layer (*SO*: 0–20 cm and 20–40 cm) as fixed effects, and also
234 their interaction ($CU \times SO$). Nutrient content of litterfall and leaf samples were analyzed by a two-way
235 ANOVA considering cutting treatment (*CT*), the taxa (*T*) and $CT \times T$ as fixed effects. Significant
236 differences were established at $p < 0.05$. To evaluate pairwise comparisons between cutting treatments
237 and taxa, the Tukey HSD test was used. All statistical analyses were performed using the R statistical
238 free software and the SPSS 19.0 software (IBM® SPSS Statistics®).

239

240 **3. Results and Discussion**

241 **3.1. Plant growth and biomass production**

242 A significant effect of the pairwise interactions $CT \times Y$ and $T \times Y$ on total dry weight (TDW)
243 was found (Table 3). TDW increased over the study period more intensively in the *III* cutting
244 treatment than in the *2I* and *3* treatments and, regarding taxa, the greatest increase in TDW
245 corresponded to *L. leucocephala* 'India' while the smallest to *L. leucocephala* 'Honduras' (Table 4,
246 Fig. 3). Analogously, WDW and NWDW showed similar results than TDW (Tables 3 and 4).
247 Therefore, both *Leucaena* species showed a good adaptation to the environmental conditions of the
248 experimental plantation as demonstrated by the high survival ($> 98.5\%$) and growth rate. The different

249 biomass components (woody, non-woody and total dry weights) were significantly correlated ($p <$
250 0.001). The correlation equation for the two species as a whole and for stems with $D > 13$ mm was
251 $WDW = 0.7768 TDW - 7.1565$ ($r = 0.99$, $n = 84$). It is noteworthy that both *Leucaena* sp. showed a
252 high fruit production after two years (up to 10.5% of the total dry weight). Interestingly, the
253 appearance of reproductive organs in *L. leucocephala* occurred just five months after planting and
254 after harvesting, and six months for *L. diversifolia*. Therefore, coppicing could enhance vegetative
255 growth in the subsequent rotation cycle by keeping the plants in a partially juvenile stage, for at least
256 5–6 months, during which the production of reproductive organs was reduced. The ratio of woody to
257 non-woody dry weight (WDW/NWDW) increased significantly from 2.5 in the $R_{1,1}S_1$ treatment, just
258 at the end of the first year, (equivalent to 71.4% WDW/TDW) to 3.0–3.5 for the rest of cutting
259 treatments (75.3–77.6% WDW/TDW).

260 [Table 3 here]

261 [Table 4 here]

262 [Figure 3 here]

263 Biomass production of the sprouts one year after cutting ($R_{2,2}S_1$, $R_{3,3}S_1$ and $R_{3,2}S_1$) was always
264 greater than that was cut every two or three years ($R_{2,1}S_2$ and $R_{3,1}S_3$), with slight but significant
265 differences between the taxa (Table 3, Figure 3). In addition, the sum of the biomass harvested
266 annually exceeded that produced by the longer two and three year rotations (Figure 3). The biomass
267 yield in our study is within the highest production range reported under favorable growing conditions
268 for *Leucaena* spp. (Parrotta, 1999; Prasad et al., 2011; Chaplot, 2014; Fagbenro et al., 2015). The first
269 year ($R_{1,1}S_1$), the biomass production of *L. leucocephala* was 72% greater than that of *L. diversifolia*,
270 even if the production of the first year for the two species was significantly lower than that of the
271 subsequent annual rotations (i.e., about 75% lower than that of $R_{2,2}S_1$ and $R_{3,3}S_1$, if measured in $Mg\ ha^{-1}$
272 ¹). This might be related to a growth slowdown during the first months of life due to planting shock
273 (Parrotta, 1992). Annual rotations ($R_{i,j}S_1$) produced shoots of a smaller diameter compared to biennial
274 or triennial rotations ($R_{2,1}S_2$ and $R_{3,1}S_3$, respectively) (Table 3, Figure 4). In spite of smaller diameter,
275 annual rotation resulted in greater biomass production per plant because of a large number of sprouts

276 (1.01 for *L. diversifolia* and 1.96 for *L. leucocephala* in the first cutting, but 4–9 in the second and 5–
277 12 thereafter for both species; only diameters greater than 15 mm were counted), in contrast to
278 biennial and triennial rotations (1.41 for *L. diversifolia* and 2.26 for *L. leucocephala* in their first
279 cutting). Consequently, coppicing reinvigorated plant growth and biomass production after each
280 harvest probably due to the breakdown of apical dominance and the subsequent development of
281 numerous new sprouts from adventitious buds that retained their juvenile stage for a few months. The
282 hormonal imbalance caused by cutting the aboveground biomass and the production of a greater
283 proportion of earlywood vessels in the new sprouts might be behind this response (Dillen et al., 2011;
284 Sjölund and Jump, 2013). It seems that coppicing *Leucaena* spp. the first year after planting produce
285 this result, and its effect was repeated in the following two years. In addition, the proliferation of
286 thinner stems in annual rotations, with higher density and less branching, facilitate the mechanization
287 of the harvest (Vanbeveren et al., 2017).

288 The annual biomass yielded in our study equals the highest yields reported in the Iberian
289 Peninsula with other lignocellulosic crops, such as kenaf, *Miscanthus* or *Eucalyptus* (Fernández et al.,
290 2018), and exceeds those reported for other tree crops, such as poplar, black locust or willow (around
291 20 Mg ha⁻¹ year⁻¹) (Sixto et al., 2015). The role of *Leucaena* as a nitrogen-fixing species, capable of
292 fixing much more than 100 kg ha⁻¹ year⁻¹ of N (Högberg and Kvarnström, 1982; Conrad et al., 2017;
293 Conrad et al., 2018), combined with their multiple uses, make it an interesting alternative short
294 rotation energy crop, at least for degraded soils and until these increase their fertility. It was reported
295 that *L. leucocephala* is capable of fixing more than 500 kg ha⁻¹ year⁻¹ of N (Duke and duCellier, 1993),
296 and it has been estimated that 73% of *L. leucocephala* N can come from atmospheric N₂-fixation
297 (Conrad et al. 2018). As a result, the nitrogen fixed can contribute to decrease the use of inorganic N-
298 fertilizers and, consequently, to reduce the high economic and energy costs related to their production
299 (Giuntoli et al., 2017). Taking into account an average TDW production of ca. 36 Mg ha⁻¹ year⁻¹
300 (annual cuttings, Table 4), and the nutrient contents in Table 5, then about 515 kg ha⁻¹ year⁻¹ of N are
301 required to support plant growth, of which 376 kg ha⁻¹ year⁻¹ could be fixed by *Leucaena-Rhizobium*
302 root nodules. As a consequence, according to Audsley et al. (2003) and Giuntoli et al. (2017), the 376
303 kg of mineral N-fertilizer saved means saving 5602 kWh ha⁻¹ year⁻¹ during manufacturing, and

304 reducing the emission of 1718 kg ha⁻¹ year⁻¹ of CO₂ into the atmosphere during manufacturing and
305 transportation (i.e. 14.9 kWh and 4.57 kg of CO₂ per kilogram of N).

306 Biomass production from the fourth to the eleventh year followed the same pattern reported
307 for the first three years. Significant effects of cutting treatment (*CT*, $p < 0.001$) and the interaction
308 between cutting treatment and year ($CT \times Y$, $p = 0.001$) when evaluating TDW from the fourth to the
309 eleventh year after planting were found (Table A.1). Annual rotation ($R_{i,j}S_1$) showed higher TDW than
310 biennial rotation ($R_{i,j}S_2$): predicted mean (SE) were 38.8 (0.3) and 22.0 (0.4), respectively (Table A.2).
311 TDW decreased with years after planting in both rotation treatments but more pronounced in biennials
312 (predicted slopes (SE) were -0.07 (0.10) and -0.63 (0.16), respectively). This trend was similar for all
313 study taxa, yet *L. leucocephala* 'India' yielded significantly more biomass ($p = 0.008$) than *L.*
314 *leucocephala* 'Honduras' (predicted mean (SE) were 32.0 (1.6) and 29.2 (1.6), respectively (Tables A.1
315 and A.2). Therefore, from the fourth to the eleventh year, annual rotations also produced more
316 biomass than biennials, so that not only environmental conditions or genotype affect production, but
317 also the harvesting frequency (Guidi Nissin et al., 2018; Kulig et al., 2019). At the end of the study
318 period, we did not observe a decrease in the vigor of sprouts, despite some stumps having up to 10
319 cuttings. The more harvests can be obtained from a vigorous stump, the longer the cultivation cycle
320 will be. This obviously delays the stump removal process and replanting costs (Dillen et al., 2011).
321 Also, we did not observe significant differences in the vigor of the sprouts between replicates or
322 between cultivation treatments.

323 Plant height was significantly influenced by the pairwise interactions $T \times CT$ and $T \times Y$ (Table
324 3). Plant height was higher in the *III* cutting treatment for *L. leucocephala* 'India' and 'K360', but in
325 the case of *L. diversifolia* height was higher in the *3* treatment compared to *III* treatment (Fig. 4). In
326 general, for *III* treatment, plant height was higher in *L. leucocephala* 'India' followed by the other
327 two *L. leucocephala* taxa with *L. diversifolia* showing lower height value, while for *3* and *2I* treatments
328 *L. diversifolia* stood out over *L. leucocephala*. In addition, during the second and third year of study,
329 height growth was significantly higher in *L. leucocephala* 'India' than in *L. diversifolia* (estimated
330 slopes (SE) were 69.1 (4.1) and 52.5 (4.1) cm year⁻¹, respectively), but *L. leucocephala* 'Honduras' and
331 'K360' (estimated slopes 65.5 (4.1) and 58.5 (4.0) cm year⁻¹, respectively) did not differ significantly

332 from the previous two taxa. Plant diameter was significantly influenced by the pairwise interactions T
333 $\times CT$ and $CT \times Y$ (Table 3). Plant diameter was significantly lower in the *111* cutting treatment for *L.*
334 *diversifolia* than in the *21* and *3* treatments. However, significant differences in plant diameter
335 between cutting treatments were not found for any of the *L. leucocephala* taxa (Fig. 4). Plant diameter
336 was generally greater in the *3* cutting treatment also showing a greater diameter increment during the
337 second and third year of study than *21* and *111* treatments (estimated slopes 6.1 (0.5), 2.3 (0.5) and 1.5
338 (0.3) cm year⁻¹, respectively).

339 [Figure 4 here]

340 Regarding phenology, both species showed sprouting and growth of buds throughout the year.
341 Maximum shoot growth was observed when maximum and minimum daily temperatures were over 20
342 °C and 12 °C, respectively. Analogously, when daily maximum and minimum temperatures were
343 below 17 °C and 5 °C, respectively, no shoot growth was reported. Fine twigs (≤ 1 cm diameter) and
344 young annual sprouts were severely affected by temperatures below -4 °C. The LAI did not display
345 significant differences between species, yet significant differences were detected for the cutting
346 treatment ($p = 0.001$) with values of 3.17 (0.09) for $R_{i,j}S_1$ and 4.00 (0.10) for $R_{2,1}S_2$ and $R_{3,1}S_3$. This
347 range of LAI values is similar to those reported for other fast growth species, such as corn or aspen,
348 since LAI is related to the exchange of energy and carbon dioxide between the crop and the
349 atmosphere, and to the biomass production (Gower et al., 1999). No significant differences between
350 taxa or cutting treatments were reported for SLA, which averaged 16.54 (0.55) m² kg⁻¹.

351 The predawn xylem water potential of the one-year-old shoots (1st and 3rd replicates) and the
352 two- or three-year-old shoots (2nd and 4th replicates) were not significantly different between taxa, yet
353 showed significant differences according to the cutting treatment. Predawn water potential in summer
354 for annual re-sprouts was always higher than -1.6 MPa, with a value of -1.1 MPa until the end of July.
355 In contrast, the two- and three-year-old shoots showed a predawn water potential of -1.5 MPa at the
356 end of July, reaching minimum values around -2.0 (0.3) MPa at the beginning of September, which
357 resulted in partial leaf abscission. *Leucaena* species are moderately tolerant to water stress (Natarajan
358 and Paliwal, 1995; Chen et al., 2012). However, at least for *Leucaena leucocephala*, when predawn
359 leaf water potential is below -1.0 MPa, gas exchange is abruptly limited, and plant survival is

360 jeopardized when it drops below -2.2 MPa (Mrema et al., 1997; Liang and Zhang, 1999). The different
361 tolerance to water stress of the annual re-sprouted shoots compared to the two- or three-year-old
362 shoots could be due to a lower soil water consumption of the harvested replicates because of a lower
363 LAI and the time required for the new sprouted shoots to reach the same size of the two- or three-year-
364 old shoots (on average it took 3–4 months and needed temperatures above 15 °C). Other experimental
365 plantations in the area under more severe rainfed conditions showed lower growth and even plant
366 mortality during summer drought, which might indicate that the present trial comprised a moderate
367 rainfed situation with a slight drought period only in summer. The low drought and cold tolerance of
368 these species probably reduces their invasive nature in the Mediterranean environment.

369

370 **3.2. Physical-chemical properties of biomass**

371 No significant differences between taxa ($p > 0.100$) were observed for dry basis heating
372 values, ash content (Table 5), wood density and bark percentage. However, we found significant
373 differences ($p < 0.05$) between the thin woody fraction (stems and branches < 25 mm of diameter) and
374 the thickest fraction (> 50 mm of diameter); mean values (SE) for wood density were 0.60 (0.03) and
375 0.74 (0.05) kg/dm³, respectively, and for bark percentage were 30.0 (2.4) and 12.0 (2.5) %,
376 respectively. The calorific capacity of both species is high enough for commercial use as an energy
377 source (Mainoo and Ulzen-Appiah, 1996; Fera et al., 2011; Fernández et al., 2015). Average annual
378 production of 25 to 35 Mg ha⁻¹ year⁻¹ considering only the woody dry biomass (excluding the first year
379 after planting), would result in 126 to 176 MWh ha⁻¹ year⁻¹, when based on the lower heating value.
380 Moreover, the non-woody dry biomass (25% of TDW) would contribute an additional 42 to 59 MWh
381 ha⁻¹ year⁻¹.

382 The mineral nutrient content of leaves, woody fraction and litterfall was not significantly
383 different between species (except for specific cases of K and Cl, Table 5). The four taxa studied
384 showed high nutrient contents, especially N, according to other studies (Parrotta, 1999; Fagbenro et
385 al., 2015) which should be taken into account in the management of plantations because of the
386 ecological value of the biogeochemical cycling of nutrients. For example, in the same way as
387 mentioned above, an annual average extraction of 29 Mg ha⁻¹ year⁻¹ of woody dry biomass (WDW)

388 implies the removal of 249, 23, 310, 160, and 64 kg ha⁻¹ year⁻¹ of N, P, K, Ca and Mg, respectively.
389 However, simply by leaving the non-woody biomass (NWDW \approx 7.25 Mg ha⁻¹ year⁻¹) in the field after
390 harvest the nutrients not removed would represent 228, 23, 106, 109 and 23 kg ha⁻¹ year⁻¹ of N, P, K,
391 Ca and Mg, respectively, which are recycled and added to the nutrients contained in the litterfall. As a
392 consequence, managers may take it into account for plantation management, both for the economic
393 and environmental aspects related to the use of inorganic fertilizers, while at the same time extracting
394 the best quality biomass for energy use (the woody biomass), since high nutrient content, mainly N, S,
395 K and Cl but also Ca and Mg, are responsible for many undesirable reactions in combustion furnaces
396 and power boilers (Jenkins et al., 1998).

397 [Table 5 here]

398 The production of pellets from the ground woody tissue had an efficiency of 99% (i.e. pellet to
399 sawdust dry weight ratio after palletization). The pellets had the following characteristics: diameter =
400 6.02 (0.01) mm, length = 19.9 (0.4) mm, bulk density = 636 (5) kg m⁻³, particle density = 1325 (15) kg
401 m⁻³ dry weight basis, mechanical durability = 96.3 (1.2) %, and moisture content = 6.1 (0.2) %.
402 However, the slightly high ash, Cl, N and S content (Table 5) could limit their application for domestic
403 furnaces of a small size, yet not for industrial uses, according to the international standards for the
404 pellets trade (TRADERSbiomass, 2020). The resulting pellets had hardwood characteristics and were
405 of better quality than pellets from non-woody species. Even so, the transformation of *Leucaena*
406 biomass into pellets confers an added value for commercialization, although the pellet quality
407 provided by debarked coniferous wood will never be achieved, as it is known that chemical
408 composition (i.e. lignin content, and the presence of hydrophobic extractives), as well as temperature
409 and pelletizing pressure, have a significant influence on the bonding quality between biomass particles
410 during the pelletizing process (Holm et al., 2006; Stelte et al., 2011). In addition, the energy use of the
411 pellets can be improved if biomass is debarked before production. For instance, the ash content of bark
412 7.6 (0.3) % was much higher than that of debarked wood, which was 1.3 (0.3) %. Moreover,
413 considering that non-woody material comprises around 25% of the total dry weight, other possible
414 uses could be considered, such as forage or composting for fertilization, given their high N, amino

415 acid and protein content (Isaac et al., 2003; Tadros et al., 2012; Tang et al., 2013; Gusha et al., 2017).
416 Nonetheless, this last issue needs further study.

417

418 **3.3. Soil properties**

419 After 11 years of short rotation cropping, on the one hand, the litter layers of the soil (O
420 horizon) showed no significant differences between taxa; the thickness of the top layer was 1.62 (0.15)
421 cm, while that of the lower stratum was 2.97 (0.25) cm, corresponding to dry weights of 1.55 (0.12) kg
422 m⁻² and 6.12 (0.66) kg m⁻², respectively. Thus, after 11 years, 77 Mg ha⁻¹ of dry matter (7.0 Mg ha⁻¹
423 year⁻¹) and 2.07 Mg ha⁻¹ of N (i.e. 188 kg ha⁻¹ year⁻¹ of N, apart from the N harvested with the
424 biomass) had accumulated in the upper part of the mineral soil. This accumulation of litterfall is much
425 greater than in forests of temperate zones (Andivia et al., 2016) and close to the amount existing in
426 tropical forests (Cassart et al., 2017), and was a consequence of the high productivity of this crop. The
427 nutrients contained in the litter represent a reservoir that will slowly be incorporated into the soil
428 (Sandhu et al., 1990; Parrotta, 1999; Gnahoua et al., 2013) which, together with the contributions from
429 the fine roots (Jha and Prasad Mohapatra, 2010), will maintain productivity of the site. The mineral
430 soil layer showed higher levels of organic matter, water-soluble organic C, N and dehydrogenase
431 activity after 11 years of a *Leucaena* biomass production system, when compared with the conditions
432 prior to establishing the system, and when compared to adjacent uncropped areas (Table 6), being
433 representative of an increase in soil fertility due to cultivation (Mrema et al., 1997; Peng et al., 2013;
434 Conrad et al., 2017). This positive effect of woody energy crops on soils have been also reported for
435 other species (Ramesh et al., 2013; Fernández et al. 2018). Therefore, leaf litter decomposition, root
436 turnover and microbial activity prevented soil degradation and, despite the continuous cropping and
437 biomass extraction, soil fertility was improved.

438 The increase of N in the soil (0–40 cm depth) compared to the uncropped areas ranged from
439 75 to 104 kg ha⁻¹ year⁻¹, which entails economic and environmental benefits, as discussed above.
440 Similarly, the organic C increase ranged from 1.2 to 2.3 Mg ha⁻¹ year⁻¹, therefore acting as a sink for
441 carbon (equivalent to 4.4 – 8.4 Mg ha⁻¹ year⁻¹ of CO₂), which allows mitigating the effects of climate
442 change. This C sequestration capacity obtained for short rotation *Leucaena* plantation is within the

443 usual high range of other plantations, 0.28–1.39 Mg ha⁻¹ year⁻¹ (Conrad et al., 2017; Campos et al.,
444 2020), although some agroforestry systems have a huge potential of C sequestration to the extent of 10
445 Mg ha⁻¹ year⁻¹ in short rotation plantations when organic amendments are added (Srinivasarao et al.,
446 2014). The importance of the increase in dehydrogenase activity is that it is a distinctive feature of
447 soil degradation metabolic pathways as it is associated with intracellular metabolism of living cells
448 and, therefore, is a good indicator of total microbial activity (Peng et al., 2013; Mukhopadhyay et al.,
449 2016). The increase in the activity of this enzyme is an unequivocal sign of the soil improvement,
450 which will lead to further growth of the plantation.

451 The other soil properties analyzed during the period of the present study (pH, EC, urease and
452 β -glucosidase activities) did not change significantly ($0.148 < p < 0.506$). The average values (SE) at
453 0–20 cm depth of pH [KCl], pH [H₂O], EC, available P and K, urease activity and β -glucosidase
454 activity for the soil treatments as a whole were 5.01 (0.16), 5.72 (0.25), 4.8 (1.5) mS m⁻¹, 13.1 (2.4)
455 mg kg⁻¹, 110 (15) mg kg⁻¹, 0.70 (0.10) μ moles g⁻¹ h⁻¹ of N-NH₄ and 51.3 (8.6) μ g g⁻¹ h⁻¹ of PNP,
456 respectively. Although no symptoms of leaf mineral deficiencies were observed throughout the study
457 period, it should be studied if some of the main nutrients, for example P and K, should be applied to
458 the soil to compensate for the exports in the harvested biomass and to prevent any reduction in
459 biomass production. In addition, we did not find any visual symptoms of soil erosion in cultivated
460 plots during the study period or plants of the cultivated taxa invading nearby areas of natural
461 vegetation.

462 [Table 6 here]

463

464 **4. Conclusions**

465 Short rotation cropping legume species of the genus *Leucaena*, harvested annually can
466 produce up to 39 Mg ha⁻¹ year⁻¹ of total dry biomass, which can generate up to 201 MWh ha⁻¹ year⁻¹
467 energy. Moreover, after 11 years of cropping, the nitrogen (≥ 35 %) and organic matter (≥ 41 %)
468 content and dehydrogenase activity (≥ 98 %) of the 0–20 cm layer of mineral soil increased compared
469 to uncropped areas and to the original soil. Therefore, soil fertility and carbon sequestration were

470 improved using short rotation cropping of *Leucaena* species. Consequently, cultivation and gains may
471 be compatible with the management of this multipurpose species in a sustainable agroforestry system.

472

473 **Appendix**

474 [Table A.1 here]

475 [Table A.2 here]

476

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482

483 **References**

484 Adhikary, P. P., Hombegowda, H. C., Barman, D., Jakhar, P., Madhu, M., 2017. Soil erosion control
485 and carbon sequestration in shifting cultivated degraded highlands of eastern India: performance
486 of two contour hedgerow systems. *Agrofor. Syst.* 91(4), 757–771.

487 <https://doi.org/10.1007/s10457-016-9958-3>.

488 Andivia, E., Rolo, V., Jonard, M., Formánek, P., Ponette, Q., 2016. Tree species identity mediates
489 mechanisms of top soil carbon sequestration in a Norway spruce and European beech mixed
490 forest. *Ann. For. Sci.* 73(2), 437–447. <https://doi.org/10.1007/s13595-015-0536-z>.

491 Audsley, E., (coord.) 2003. Harmonisation of environmental life cycle assessment for agriculture,
492 Final Report. Concerted Action AIR3-CT94-2028, European Commission, Directorate-General
493 VI Agriculture, Brussels, 2003, p. 129 B-1049 Belgium.

494 Beck, H. E., Zimmermann, N.E., McVicar, T.R., Vergopolan, N., Berg, A., Wood, E.F., 2018. Present
495 and future Köppen-Geiger climate classification maps at 1-km resolution. *Sci. Data.* 5:180214

496 <https://doi.org/10.1038/sdata.2018.214>.

497 Bogdanski, A., Dubois, O., Jamieson, C., Krell, R., 2010. Making integrated food-energy systems
498 work for people and climate. FAO, Roma. 116 p.

499 Bray, R.A., Sorensson, C.T., 1992. *Leucaena diversifolia*, fast growing highland NFT species. In:
500 Nitrogen Fixing Trees Highlights, NFTA 92-05. Nitrogen Fixing Tree Association, USA.

501 Cassart, B., Basia, A.A., Titeux, H., Andivia, E., Ponette, Q., 2017. Contrasting patterns of carbon
502 sequestration between *Gilbertiodendron dewevrei* monodominant forests and *Scorodophloeus*
503 *zenkeri* mixed forests in the Central Congo basin. Plant Soil 414(1–2), 309–326.
504 <https://doi.org/10.1007/s11104-016-3130-8>.

505 Campos, R., Pires, G.F., Costa, M.H., 2020. Soil carbon sequestration in rainfed and irrigated
506 production systems in a New Brazilian agricultural frontier. Agriculture 10(5), 156.
507 <https://doi.org/10.3390/agriculture10050156>.

508 Chaplot, P.C., 2014. Effect of planting geometry on biomass partitioning and productivity of *Leucaena*
509 *leucocephala*. Ann. Agri Bio Res. 19(4), 703–705.

510 Chen, Y., Chen, F., Liu, L., Zhu, S., 2012. Physiological responses of *Leucaena leucocephala*
511 seedlings to drought stress. Procedia Eng. 28: 110–116.
512 <https://doi.org/10.1016/j.proeng.2012.01.691>.

513 Chiou, C.-R., Chen, Y.-J., Wang, H.-H., Grant, W.E., 2016. Predicted range expansion of the invasive
514 plant *Leucaena leucocephala* in the Hengchun peninsula, Taiwan. Biol. Invasions 18, 381–394.
515 <https://doi.org/10.1007/s10530-015-1010-4>.

516 Christersson, L., Verma, K., 2006. Short-rotation forestry – a complement to “conventional” forestry.
517 Unasylva, n° 223, vol. 57: 34-39. FAO, Roma. (available at:
518 <http://www.fao.org/docrep/pdf/008/A0532e/A0532e07.pdf>)

519 Conrad, K.A., Dalal, R.C., Dalzell, S.A., Allen, D.E., Fujinuma, R., Menzies, N.W., 2018. Soil
520 nitrogen status and turnover in subtropical leucaena-grass pastures as quantified by delta ¹⁵N
521 natural abundance. Geoderma 313, 126–134. <https://doi.org/10.1016/j.geoderma.2017.10.029>.

522 Conrad, K.A., Dalal, R.C., Dalzell, S.A., Allen, D.E., Menzies, N.W., 2017. The sequestration and
523 turnover of soil organic carbon in subtropical leucaena-grass pastures. *Agric. Ecosyst. Environ.*
524 248, 38–47. <https://doi.org/10.1016/j.agee.2017.07.020>.

525 Dalzell, S. A., 2019. *Leucaena* cultivars – current releases and future opportunities. *Tropical*
526 *Grasslands-Forrajés Tropicales* 7(2), 56–64. [https://doi.org/10.17138/TGFT\(7\)56-64](https://doi.org/10.17138/TGFT(7)56-64).

527 Dillen, S.Y., El Kasmioui, O., Marron, N., Calfapietra, C., Ceulemans, R., 2011. Poplar. In: Halford,
528 N.G., Karp, A. (Eds.), *Energy crops*. RSC Energy and Environment Series n°3. RSC Publishing,
529 Cambridge, UK. pp. 275-300. <https://doi.org/10.1039/9781849732048>.

530 Duke, J.A., duCellier, J.L.. 1993. *CRC Handbook of alternative cash crops*. CRC Press, Boca Ratón,
531 Florida, USA. 544 p. ISBN: 978-0-8493-3620-1.

532 European Union, 2018. Directive (EU) 2018/2001 of the European Parliament and of the Council of
533 11 December 2018 on the promotion of the use of energy from renewable sources. (available at:
534 <https://eur-lex.europa.eu/legal-content/EN/TXT/PDF/?uri=CELEX:32018L2001&from=EN>

535 Fagbenro, J.A., Oshunsanya, S.O., Aluko, P.A., Oyeleye, B.A., 2015. Biomass production, tissue
536 nutrient concentration, and N₂-fixing potentials of seven tropical leguminous species. *Commun.*
537 *Soil Sci. Plant Anal.* 46(6), 709–723. <https://doi.org/10.1080/00103624.2015.1005221>.

538 FAO, 2015. *Combating land degradation for food security and provision of soil ecosystem services in*
539 *Europe and Central Asia – International Year of Soil 2015*. Ed. Food and Agriculture
540 Organization of United Nations (FAO). 39th session of the European Commission on Agriculture
541 Budapest, Hungary, 22-23 September 2015. Report ECA/39/13/3. (available at:
542 <http://www.fao.org/europe/commissions/eca/eca-39/en/>)

543 Feria, M.J., López, F., García, J.C., Pérez, A., Zamudio, M.A.M., Alfaro, A., 2011. Valorization of
544 *Leucaena leucocephala* for energy and chemicals from autohydrolysis. *Biomass Bioenergy* 35,
545 2224–2233. <http://doi.org/10.1016/j.biombioe.2011.02.038>.

546 Fernández, M., Alaejos, J., Andivia, E., Vázquez-Piqué, J., Ruiz, F., López, F., Tapias, R., 2018.
547 *Eucalyptus x urograndis* biomass production for energy purposes exposed to a Mediterranean

548 climate under different irrigation and fertilisation regimes. *Biomass Bioenergy* 111, 22–30.
549 <https://doi.org/10.1016/j.biombioe.2018.01.020>.

550 Fernández, M., García-Albalá, J., Andivia, E., Alaejos, J., Tapias, R., Menéndez, J., 2015. Sickie bush
551 (*Dichrostachys cinerea* L.) field performance and physicochemical property assessment for
552 energy purposes. *Biomass Bioenergy* 81, 483–489.
553 <https://doi.org/10.1016/j.biombioe.2015.08.006>.

554 Foroughbakhch, R., Hernández-Piñero, J.L., Ramírez, R., Alvarado, M.A., 2007. Nutrients, mineral
555 and volatile fatty acids content in four *Leucaena* species and the hybrid K743. *J. Anim. Vet. Adv.*
556 6(9), 1083–1087.

557 Giuntoli, J., Agostini, A., Edwards, R., Marelli, L., 2017. Solid and gaseous bioenergy pathways: input
558 values and GHG emissions. Calculated according to the methodology set in COM(2016) 767,
559 EUR 27215 EN, doi:10.2790/27486.

560 Gnahoua, G.M., Oliver, R., Nguessan, K.A., Balle, P., 2013. Production and mineral impact of litter
561 from leguminous trees used to improve fallow land in the forest zone of Cote d'Ivoire. [in
562 French]. *Journal of Applied Biosciences* 72, 5800-5809. <https://doi.org/10.4314/jab.v72i1.99665>.

563 Gower, S.T., Kucharik, C.J., Norman, J.M., 1999. Direct and indirect estimation of leaf area index,
564 f_{APAR} , and net primary production of terrestrial ecosystems. *Remote Sens. Environ.* 70, 29–51.

565 Guidin Nissin, W., Lafleur, B., Labrecque, M., 2018. The performance of five willow cultivars under
566 different pedoclimatic conditions and rotation cycles. *Forests* 9(6), 349.
567 <https://doi.org/10.3390/f9060349>.

568 Gusha, J., Chiuta, T., Katsande, S., Zvinorova, P.I., Kagande, S.M., 2017. Performance of cattle reared
569 on rangelands supplemented with farm-formulated diets during the dry season in Zimbabwe.
570 *Anim. Prod. Sci.* 57(6), 1163–1169. <https://doi.org/10.1071/AN15900>.

571 Halford, N.G., Karp, A. (Eds.), 2011. Energy crops. RSC Energy and Environment Series n°3. RSC
572 Publishing, Cambridge, UK. 442 p. <https://doi.org/10.1039/9781849732048>.

573 Högberg, P., Kvarnström, M., 1982. Nitrogen fixation by the woody legume *Leucaena leucocephala*
574 in Tanzania. *Plant Soil* 66, 21–28.

575 Holm, J.K., Henriksen, U.B., Hustad, J.E., Sørensen, L.H., 2006. Toward an understanding of
576 controlling parameters in softwood and hardwood pellets production. *Energ. Fuel.* 20,
577 2686–2694.

578 IEA, 2016. International Energy Agency, World energy outlook 2016, Paris, France, 2016, 684 p.
579 ISBN: 978-92-64-26494-6.

580 Isaac, L., Wood, C.W., Shannon, D.A., 2003. Pruning management effects on soil carbon and nitrogen
581 in contour-hedgerow cropping with *Leucaena leucocephala* (Lam.) De Wit on sloping land in
582 Haiti. *Nutr. Cycl. Agroecosys.* 65, 253–263. <https://doi.org/10.1023/A:1022600720226>.

583 Jenkins, B.M., Baxter, L.L., Miles Jr., T.R., Miles, T.R., 1998. Combustion properties of biomass.
584 *Fuel Process. Technol.* 54, 17–46.

585 Jha, P., Prasad Mohapatra, K., 2010. Leaf litterfall, fine root production and turnover in four major
586 tree species of the semi-arid region of India. *Plant Soil* 326, 481–491.
587 <https://doi.org/10.1007/s11104-009-0027-9>.

588 Katunga, M.M.D., Muhigwa, B.J.B., Kashala, K.J.C., Kambuyi, M., Nyongombe, N., Maass, B.L.,
589 Peters, M., 2014. Agro-ecological adaptation and participatory evaluation of multipurpose tree
590 and shrub legumes in mid altitudes of Sud-Kivu, D. R. Congo. *Am. J. Plant Sci.* 5, 2031–2039.
591 <https://doi.org/10.4236/ajps.2014.513218>.

592 Kulig, B., Gacek, E., Wojciechowski, R., Oleksy, A., Kołodziejczyk, M., Szewczyk, W., Klimek-
593 Kopyra, A., 2019. Biomass yield and energy efficiency of willow depending on cultivar,
594 harvesting frequency and planting density. *Plant Soil Environ.*, 65: 377–386.
595 <https://doi.org/10.17221/594/2018-PSE>.

596 Lauri, P., Havlík, P., Kindermann, G., Forsell, N., Böttcher, H., Obersteiner, M., 2014. Woody
597 biomass energy potential in 2050. *Energy Policy* 66, 19–31.
598 <https://doi.org/10.1016/j.enpol.2013.11.033>.

599 Leketa, K., Donkin, E. F., Hassen, A., Akanmu, A. M., 2019. Effect of *Leucaena leucocephala*, as a
600 protein source in a total mixed ration, on milk yield and composition of Saanen milk goats. *S.*
601 *Afr. J. Anim. Sci.* 49(2), 301–309. <http://dx.doi.org/10.4314/sajas.v49i2.10>.

602 Liang, J., Zhang, J., 1999. The relations of stomatal closure and reopening to xylem ABA
603 concentration and leaf water potential during soil drying and rewatering. *Plant Growth Regul.* 29,
604 77–86. <https://doi.org/10.1023/A:1006207900619>.

605 Loaiza, J.M., López, F., García, M.T., García, J.C., Díaz, M.J., 2017. Biomass valorization by using a
606 sequence of acid hydrolysis and pyrolysis processes. Application to *Leucaena leucocephala*. *Fuel*
607 203, 393–402. <http://dx.doi.org/10.1016/j.fuel.2017.04.135>.

608 Mainoo, A.A., Ulzen-Appiah, F., 1996. Growth, wood yield and energy characteristics of *Leucaena*
609 *leucocephala*, *Gliricidia sepium* and *Senna siamea* at age four years. *Ghana Journal of Forestry*
610 13, 69–79.

611 Malik, R.S., Dutt, D., Tyagi, C.H., Jindal, A.K., Lakharia, L.K., 2004. Morphological, anatomical and
612 chemical characteristics of *Leucaena leucocephala* and its impact on pulp and paper making
613 properties. *J. Sci. Ind. Res.* 63(2), 125–133.

614 Mrema, A.F., Granhall, U., Sennerby-Forsse, L., 1997. Plant growth, leaf water potential, nitrogenase
615 activity and nodule anatomy in *Leucaena leucocephala* as affected by water stress and nitrogen
616 availability. *Trees* 112, 42–48. <https://doi.org/10.1007/PL00009695>.

617 Mukhopadhyay, S., Mastro, R.E., Yadav, A., George, J., Ram, L.C., Shukla, S.P., 2016. Soil quality
618 index for evaluation of reclaimed coal mine spoil. *Sci. Total Environ.* 542, 540–550.
619 <https://doi.org/10.1016/j.scitotenv.2015.10.035>.

620 Nakagawa, S., Schielzeth, H., 2013. A general and simple method for obtaining R^2 from generalized
621 linear mixed effects models. *Methods Ecol. Evol.* 4, 132–142. [https://doi.org/10.1111/j.2041-](https://doi.org/10.1111/j.2041-210x.2012.00261.x)
622 [210x.2012.00261.x](https://doi.org/10.1111/j.2041-210x.2012.00261.x)

623 Natarajan, K., Paliwal, K., 1995. Photosynthesis by *Leucaena leucocephala* during water stress.
624 Nitrogen Fixing tree Research Reports 13, 79-83. (available at:
625 [http://www.nzdl.org/gsdmod?e=d-00000-00---off-0hdl--00-0----0-10-0---0---0direct-10---4-----](http://www.nzdl.org/gsdmod?e=d-00000-00---off-0hdl--00-0----0-10-0---0---0direct-10---4-----0-11--11-en-50---20-about---00-0-1-00-0--4----0-0-11-10-0utfZz-8-00&a=d&c=hdl&cl=CL1.12&d=HASH01dc98a04f289ccbe1563ac7.2.19#HASH01dc98a04f289ccbe1563ac7.2.19)
626 [0-11--11-en-50---20-about---00-0-1-00-0--4----0-0-11-10-0utfZz-8-](http://www.nzdl.org/gsdmod?e=d-00000-00---off-0hdl--00-0----0-10-0---0---0direct-10---4-----0-11--11-en-50---20-about---00-0-1-00-0--4----0-0-11-10-0utfZz-8-00&a=d&c=hdl&cl=CL1.12&d=HASH01dc98a04f289ccbe1563ac7.2.19#HASH01dc98a04f289ccbe1563ac7.2.19)
627 [00&a=d&c=hdl&cl=CL1.12&d=HASH01dc98a04f289ccbe1563ac7.2.19#HASH01dc98a04f289](http://www.nzdl.org/gsdmod?e=d-00000-00---off-0hdl--00-0----0-10-0---0---0direct-10---4-----0-11--11-en-50---20-about---00-0-1-00-0--4----0-0-11-10-0utfZz-8-00&a=d&c=hdl&cl=CL1.12&d=HASH01dc98a04f289ccbe1563ac7.2.19#HASH01dc98a04f289ccbe1563ac7.2.19)
628 [ccbe1563ac7.2.19](http://www.nzdl.org/gsdmod?e=d-00000-00---off-0hdl--00-0----0-10-0---0---0direct-10---4-----0-11--11-en-50---20-about---00-0-1-00-0--4----0-0-11-10-0utfZz-8-00&a=d&c=hdl&cl=CL1.12&d=HASH01dc98a04f289ccbe1563ac7.2.19#HASH01dc98a04f289ccbe1563ac7.2.19))

629 Noulekoun, F., Naab, J.B., Lamers, J.P.A., Baumert, S., Khamzina, A., 2018. Sapling biomass
630 allometry and carbon content in five afforestation species on marginal farmland in semi-arid
631 Benin. *New For.* 49(3), 363–382. <https://doi.org/10.1007/s11056-017-9624-2>.

632 Mello, T.J., Oliveira, A.A., 2016. Making a bad situation worse: an invasive species altering the
633 balance of interactions between local species. *PLoS ONE*, 11(3): e0152070.
634 <https://doi.org/10.1371/journal.pone.0152070>.

635 Parrotta, J.A., 1992. *Leucaena leucocephala* (Lam.) de Wit: leucaena, tantan. Res. Note SO-ITFSM-
636 52. New Orleans, LA. USDA Forest Service. Southern Forest Experiment Station, 8 p.

637 Parrotta, J.A., 1999. Productivity, nutrient cycling, and succession in single- and mixed-species
638 plantations of *Casuarina equisetifolia*, *Eucalyptus robusta*, and *Leucaena leucocephala* in Puerto
639 Rico. *For. Ecol. Manage.* 124, 45–77. [https://doi.org/10.1016/S0378-1127\(99\)00049-3](https://doi.org/10.1016/S0378-1127(99)00049-3).

640 Peng, S., Chen, A., Fang, H., Wu, J., Liu, G., 2013. Effects of vegetation restoration types on soil
641 quality in Yuanmou dry-hot valley, China. *Soil Sci. Plant Nutr.* 59(3), 347–360.
642 <https://doi.org/10.1080/00380768.2013.785918>.

643 Pinheiro, J., Bates, D., DebRoy, S., Sarkar, D., Core Team, R., 2019. nlme: Linear and Nonlinear
644 Mixed Effects Models. R package version 3.1-142. (available at: [https://CRAN.R-](https://CRAN.R-project.org/package=nlme)
645 [project.org/package=nlme](https://CRAN.R-project.org/package=nlme))

646 Prasad, J.V.N.S., Korwar, G.R., Rao, K.V., Mandal, U.K., Rao, G.R., Srinivas, I., Venkateswarlu, B.,
647 Rao, S.N., Kulkarni, H.D., 2011. Optimum stand density of *Leucaena leucocephala* for wood
648 production in Andhra Pradesh, Southern India. *Biomass Bioenergy* 35, 227–235.
649 <https://doi.org/10.1016/j.biombioe.2010.08.012>.

650 Ramesh, T., Manjaiah, K.M., Tomar, J.M.S., Ngachan, S.V., 2013. Effect of multipurpose tree
651 species on soil fertility and CO₂ efflux under hilly ecosystems of Northeast India. *Agrofor. Syst.*
652 87, 1377–1388. <https://doi.org/10.1007/s10457-013-9645-6>.

653 REN21, 2017. Renewables 2017 Global Status Report, Paris, France. 301 p. ISBN: 978-3-9818107-6-
654 9. (available at: [http://www.ren21.net/wp-content/uploads/2017/06/17-](http://www.ren21.net/wp-content/uploads/2017/06/17-8399_GSR_2017_Full_Report_0621_Opt.pdf)
655 [8399_GSR_2017_Full_Report_0621_Opt.pdf](http://www.ren21.net/wp-content/uploads/2017/06/17-8399_GSR_2017_Full_Report_0621_Opt.pdf))

656 Sa'ad, M.F., Yunus, N.Y., Ab Rahman, H., Wan Abdul Rahman, W.M.N., 2019. *Leucaena*
657 particleboard: a commercial trial. *BioResources* 14(2), 3506–3511.

658 Sandhu, J., Sinha, M., Ambash, R.S., 1990. Nitrogen release from decomposing litter of *Leucaena*
659 *leucocephala* in the dry tropics. *Soil Bid. Biochem.* 22(6), 859–863.
660 [https://doi.org/10.1016/0038-0717\(90\)90168-Y](https://doi.org/10.1016/0038-0717(90)90168-Y).

661 Sangram, C., Keerthika, A., 2013. Genetic Variability and association studies among morphological
662 traits of *Leucaena leucocephala* (Lam.) de Wit. genetic resources. *Res. J. Agriculture and*
663 *Forestry Sci.* 1(8), 23–29.

664 Santana, A. A., Cheng, L., Verdecia, D. M., Ramírez, J. L., López, S., Cisneros, M. V., Rugoho, I.,
665 Maxwell, T. M. R., Al-Marashdeh, O., 2019. Effect of a mixed silage of king grass (*Cenchrus*
666 *purpureus*) and forage legumes (*Leucaena leucocephala* or *Gliricidia sepium*) on sheep intake,
667 digestibility and nitrogen balance. *Anim. Prod. Sci.* 59(12), 2259–2264.
668 <https://doi.org/10.1071/AN18559>.

669 Sixto, H., Cañellas, I., Arendonk, J., Ciria, P., Camps, F., Sánchez, M., Sánchez-González, M., 2015.
670 Growth potential of different species and genotypes for biomass production in short rotation in
671 Mediterranean environments. *For. Ecol. Manage.* 354, 291–299.
672 <https://doi.org/10.1016/j.foreco.2015.05.038>.

673 Shelton, M., Dalzell, S., 2007. Production, economic and environmental benefits of *Leucaena*
674 pastures. *Tropical Grasslands* 41, 174–190.

675 Sjölund, M.J., Jump, A.S., 2013. The benefits and hazards of exploiting vegetative regeneration for
676 forest conservation management in a warming world. *Forestry* 86, 503–513.
677 <https://doi.org/10.1093/forestry/cpt030>

678 Srinivasarao, C., Lal, R., Sumanta Kundu Babu, M.B.B.P., Venkateswarlu, B., Singh, A.K., 2014.
679 Soil carbon sequestration in rainfed production systems in the semiarid tropics of India. *Sci. Total*
680 *Environ.* 487, 587–603. <https://doi.org/10.1016/j.scitotenv.2013.10.006>.

681 Stelte, W., Holm, J.K., Sanadi, A.R., Barsberg, S., Ahrenfeldt, J., Henriksen, U.B., 2011. A study of
682 bonding and failure mechanisms in fuel pellets from different biomass resources. *Biomass*
683 *Bioenergy* 35(2), 910–918. <https://doi.org/10.1016/j.biombioe.2010.11.003>.

684 Tabatabai, M.A., 1994. Soil enzymes. In: Weaver, R.W., Angle, S., Bottomley, P. (Eds.), *Methods of*
685 *Soil Analysis. Part 2: Microbiological and Biochemical Properties*. Soil Science Society of
686 America, Madison, WI (USA), pp. 775–833.

687 Tadros, M.J., Al-Mefleh, N.K., Chandler, P., 2012. Morphology, productivity and forage quality of
688 *Leucaena leucocephala* as influenced by irrigation under field conditions. *Agrofor. Syst.* 86, 73–
689 81. <https://doi.org/10.1007/s10457-012-9539-z>.

690 Tang, G., Li, K., Zhang, C., Gao, C., Li, B., 2013. Accelerated nutrient cycling via leaf litter, and not
691 root interaction, increases growth of *Eucalyptus* in mixed-species plantations with *Leucaena*. *For.*
692 *Ecol. Manage.* 310, 45–53. <https://doi.org/10.1016/j.foreco.2013.08.021>.

693 TRADERSbiomass, 2020. [https://tradersbiomass.com/online-biomass-exchangetrading-with-](https://tradersbiomass.com/online-biomass-exchangetrading-with-tradersbiomass/wood-pellet-trading/)
694 [tradersbiomass/wood-pellet-trading/](https://tradersbiomass.com/online-biomass-exchangetrading-with-tradersbiomass/wood-pellet-trading/), (accessed 04 June 2020).

695 United Nations, 2016. Report of the Conference of the Parties on its twenty-first session, held in Paris
696 from 30 November to 13 December 2015 (FCCC/CP/2015/10/Add.1), (2016).
697 <http://unfccc.int/resource/docs/2015/cop21/eng/10a01.pdf>. (Accessed 06 June 2020).

698 Vanbeveren, S.P.P., Schweier, J., Berhongaray, G., Ceulemans, R., 2015. Operational short rotation
699 woody crop plantations: Manual or mechanised harvesting? *Biomass Bioenergy* 72, 8–18.
700 <http://dx.doi.org/10.1016/j.biombioe.2014.11.019>.

701 Vanguelova, E., Pitman, R., 2011. Impacts of short rotation forestry on soil sustainability. In: McKay,
702 H. (Ed.), *Short rotation forestry: review of growth and environmental impacts*, Forest Research
703 Monograph, 2, Forest Research, Surrey, UK, pp. 37–78. ISBN 978-0-85538-827-0.

704 Zayed, M.Z., Sallam, S.M.A., Shetta, N.D., 2018. Review article on *Leucaena leucocephala* as one of
705 the miracle timber trees. *Int. J. Pharm. Pharm. Sci.*, 10 (1), 1–7.
706 <http://dx.doi.org/10.22159/ijpps.2018v10i1.18250>.

707 Zayed, M.Z., Wu, A., Sallam, S., 2019. Comparative phytochemical constituents of *Leucaena*
708 *leucocephala* (Lam.) leaves, fruits, stem barks, and wood branches grown in Egypt using GC-MS
709 method coupled with multivariate statistical approaches. BioResources 14(1), 996–1013.
710 Zuur, A. F., Ieno, E.N., Walker, N., Saveliev, A.A., Smith, G.M., 2009. Mixed effects models and
711 extensions in ecology with R. Springer, New York, NY, USA. ISBN 978-0-387-87458-6
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714 Table 1: Climatic conditions in the area during the study period.

Climatic variable	Mean (standard error)
Absolute maximum temperature (°C)	43.8
Warmest month's average maximum temperature (°C)	31.3 (0.8)
Average annual temperature (°C)	16.7 (0.3)
Coldest month's average minimum temperature (°C)	4.5 (1.7)
Absolute minimum temperature (°C)	-5.4
Annual precipitation (mm)	556 (197)
Summer precipitation from June to August (mm)	16 (11)
Annual reference crop evapotranspiration, ETo (mm)	1109 (30)

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719 Table 2: Physical-chemical properties of litterfall, soil and plant material determined, together with brief
 720 explanation of the technical methods and instruments used.

Sample	Property	Technical method	Observations
Litterfall, soil and plant	Total N	Kjedahl method (auto-analyser Bran+Luebbe®, Mod. AIII).	
Litterfall and plant	C, S	Elemental analyzer (Thermo Scientific™ FLASH 2000).	
Litterfall, soil and plant	Ca, P, K, Mg	ICP-OES (Thermo Jarrell Ash Corporation). Olsen method for available P in soil.	After undertaking the total extraction with HNO ₃ . Available K in soil after extraction with ammonium acetate.
Soil	pH, Electrical conductivity (EC)	Multiparameter bench (PC 80, XS® Instruments)	pH: with distilled water (pH [H ₂ O]) and also in basis of a KCl solution (pH [KCl]). Volume fraction 1:2.5 for pH and 1:5 for EC.
Soil	Oxidizable organic matter fraction (OM)	Walkley and Black method.	
Soil	Water-soluble carbon (WC)	Combustion catalytic oxidation (Carbon TOC-VCSH, Shimadzu®).	By extraction with Milli-Q water following a 1:10 (soil to water) ratio.
Soil (0-20 cm layer)	β-glucosidase activity (β-gluc), dehydrogenase activity (Dhe), urease activity (Urease)	Standardized methods for these biochemical properties (Tabatabai, 1994).	β-gluc (p-nitrophenol content). Dhe (iodonitrotetrazolium formazan content). Urease (N-NH ₄ content).
Plant	Cl	Photometer (HI 83200, Hanna® Instruments).	After undertaking the total extraction with HNO ₃ .
Plant	Ash	Muffle oven	550 °C (ISO 18122:2015).
Plant	Heating values (constant volume)	According to the European standard EN 14918:2011 for solid biofuels using an automatic isoperibol calorimeter (Parr 6300®)	- Cold milling in order to avoid chemical structure disruptions. - Crushed to a powdered form and sieved through a 1 mm size mesh. - Data referred to a dry basis (moisture-free) after oven-drying at 105 °C (ISO 18134-3:2015).

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726 Table 3: *F* and *p* values (in parentheses) for the fixed effects of the linear mixed models to evaluate height (*H*),
 727 stem diameter (*D*), as well as total (TDW), woody (WDW) and non-woody (NWDW) dry weight, over the first
 728 three years after planting. Marginal and conditional R^2 for each model is also shown. Effects: *Y*, year; *CT*,
 729 cutting treatment; *T*, taxa.

	<i>H</i> (cm)	<i>D</i> (mm)	TDW (Mg ha ⁻¹)	WDW (Mg ha ⁻¹)	NWDW (Mg ha ⁻¹)
<i>Y</i>	224.0 (<0.001)	152.4 (<0.001)	224.0 (<0.001)	227.5 (<0.001)	198.3 (<0.001)
<i>CT</i>	5.1 (0.080)	10.5 (0.026)	9.3 (0.031)	4.5 (0.095)	7.0 (0.049)
<i>T</i>	7.7 (<0.001)	4.6 (0.003)	4.4 (0.020)	3.4 (0.046)	5.9 (0.007)
<i>CT</i> × <i>Y</i>	0.4 (0.633)	32.6 (<0.001)	6.7 (0.002)	4.2 (0.021)	14.6 (<0.001)
<i>T</i> × <i>Y</i>	3.4 (0.018)	0.7 (0.521)	3.3 (0.026)	3.4 (0.025)	3.0 (0.040)
<i>CT</i> × <i>T</i>	16.8 (<0.001)	10.7 (<0.001)	1.3 (0.310)	0.8 (0.560)	2.0 (0.126)
<i>CT</i> × <i>T</i> × <i>Y</i>	1.2 (0.395)	1.1 (0.421)	0.9 (0.483)	1.0 (0.457)	0.8 (0.576)
<i>R</i> ² -marginal	0.566	0.324	0.591	0.576	0.573
<i>R</i> ² -conditional	0.579	0.426	0.717	0.712	0.720

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734 Table 4: Estimated year slope [mean (standard error)] for each cutting treatment and taxa of the linear mixed
 735 models to evaluate accumulated total (TDW), woody (WDW) and non-woody (NWDW) dry weight over the
 736 first three years after planting. Different letters depict significant differences ($p < 0.05$) between cutting
 737 treatments and taxa, respectively. Cutting treatments: *111* ($R_{11}S_1 + R_{22}S_1 + R_{33}S_1$), *21* ($R_{21}S_2 + R_{32}S_1$) and *3*
 738 ($R_{31}S_3$). Taxa: *Ld*, *Leucaena diversifolia*; *Ll*, *Leucaena leucocephala*.

	TDW	WDW	NWDW
	(Mg ha ⁻¹ year ⁻¹)	(Mg ha ⁻¹ year ⁻¹)	(Mg ha ⁻¹ year ⁻¹)
Cutting treatment			
<i>111</i>	35.7 (1.3) a	25.0 (0.9) a	9.8 (0.4) a
<i>21</i>	21.1 (1.8) b	15.6 (1.3) b	5.4 (0.5) b
<i>3</i>	10.1 (1.8) c	7.9 (1.3) c	2.1 (0.5) c
Taxa			
<i>Ld</i>	22.2 (1.9) b	15.8 (1.3) b	6.0 (0.6) b
<i>Ll</i> 'Honduras'	15.3 (1.9) c	10.8 (1.3) c	4.1 (0.6) c
<i>Ll</i> 'India'	29.7 (1.9) a	20.9 (1.3) a	8.4 (0.6) a
<i>Ll</i> 'K360'	22.0 (1.9) bc	15.8 (1.3) b	5.9 (0.6) bc

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744 Table 5: Chemical and energetic properties [mean (SE)], expressed on a dry weight basis, for the different
 745 samples (leaves, woody tissues [stem and branches with bark] and litterfall) for the four taxa of *Leucaena* as a
 746 whole.

	Leaves	Woody tissues	Litterfall
C (%)	44.94 (0.52)	44.98 (0.78)	42.63 (0.84)
N (%)	3.14 (0.19)	0.86 (0.12)	2.67 (0.25)
P (%)	0.32 (0.05)	0.08 (0.02)	0.13 (0.02)
K (%)	1.47 (0.12)	1.07 (0.08)	0.32 (0.11)*
Ca (%)	1.50 (0.16)	0.55 (0.20)	1.47 (0.44)
Mg (%)	0.32 (0.02)	0.22 (0.02)	0.20 (0.01)
S (%)	0.28 (0.05)	0.09 (0.03)	0.45 (0.05)
Cl (%)	0.27 (0.03)*	0.08 (0.02)	<i>n.a.</i>
HHV (MJ/kg)	19.7 (0.2)	19.4 (0.2)	<i>n.a.</i>
LHV (MJ/kg)	18.4 (0.2)	18.1 (0.2)	<i>n.a.</i>
Ash (%)	10.5 (0.3)	1.89 (0.03)	<i>n.a.</i>

747 Asterisks indicate significant differences ($p < 0.05$) among taxa, with *L. diversifolia* (*Ld*) being significantly
 748 different from the three taxa of *L. leucocephala* (*Ll*): K = 0.45 (0.08) % for *Ld*, and K = 0.18 (0.07) % for *Ll* as a
 749 whole; Cl = 0.23 (0.02) % for *Ld*, and Cl = 0.33 (0.02) % for *Ll* as a whole. *n.a.*: not analyzed.

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754 Table 6: Soil characteristics [mean (SE)] at two depths (*SOD*, 0–20 cm and 20–40 cm). Different letters within a
 755 row denote significant differences ($p < 0.05$) between cultivation treatments (*CU*). OM: organic matter; WC:
 756 water-soluble carbon; Dhe: dehydrogenase activity ($\mu\text{g g}^{-1} \text{h}^{-1}$ of INTF).

<i>Parameter</i>	<i>Soil Depth (SOD)</i>	<i>Cultivation treatment (CU)</i> ⁽¹⁾			<i>Significance level (p)</i>		
		<i>Uncultivated</i>	<i>Ld</i>	<i>Ll</i> ⁽²⁾	<i>CU</i>	<i>SOD</i>	<i>CU x SOD</i>
OM (%)	<i>0-20 cm</i>	1.03 (0.12) a	1.85 (0.06) b	1.45 (0.21) b	< 0.001	0.110	0.866
	<i>20-40 cm</i>	0.66 (0.09) a	1.70 (0.15) b	1.21 (0.18) b			
WC (mg kg ⁻¹)	<i>0-20 cm</i>	139 (20) a	186 (13) b	172 (22) ab	0.004	0.745	0.104
	<i>20-40 cm</i>	112 (22)	146 (10)	117 (16)			
N (mg kg ⁻¹)	<i>0-20 cm</i>	390 (76) a	625 (50) b	525 (77) ab	0.012	0.060	0.787
	<i>20-40 cm</i>	230 (32) a	493 (20) b	455 (68) b			
Dhe	<i>0-20 cm</i>	0.66 (0.04) a	1.40 (0.21) b	1.31 (0.10) b	0.015		

757 ⁽¹⁾ Five cultivation treatments were used (*CU*): uncultivated (natural vegetation), cultivated under *L. diversifolia*
 758 (*Ld*), and three treatments cultivated under each *L. leucocephala* taxon (*Ll*).

759 ⁽²⁾ As *L. leucocephala* taxa did not differ significantly between each other, the mean values for the three taxa as
 760 a whole are shown.

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764 Table A.1: *F* and *p* values (in parentheses) for the fixed effects of the linear mixed models to evaluate total dry
765 weight (TDW) for the period between the fourth and the eleventh year after planting. Marginal and conditional
766 R^2 for each model is also shown. Effects: *Y*, year; *CT*, cutting treatment; *T*, taxa

	TDW (Mg ha ⁻¹ year ⁻¹)
<i>Y</i>	3.7 (0.059)
<i>CT</i>	1203.3 (<0.001)
<i>T</i>	5.0 (0.008)
<i>CT</i> × <i>Y</i>	13.0 (0.001)
<i>T</i> × <i>Y</i>	0.2 (0.995)
<i>CT</i> × <i>T</i>	0.2 (0.929)
<i>CT</i> × <i>T</i> × <i>Y</i>	0.1 (0.998)
<i>R</i> ² -marginal	0.907
<i>R</i> ² -conditional	0.907

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770 Table A.2: Estimated total dry biomass [mean (standard error)] for each cutting treatment and taxa evaluated by
 771 the linear mixed models from the fourth to the eleventh year after planting. Different letters depict significant
 772 differences ($p < 0.05$) between cutting treatments and taxa, respectively. Cutting treatments: *annual* (R_{ij}S₁) and
 773 biennial (R_{ij}S₂). Taxa: *Ld*, *Leucaena diversifolia*; *Ll*, *Leucaena leucocephala*.

TDW	
(Mg ha ⁻¹ year ⁻¹)	
Cutting treatment	
<i>annual</i>	38.8 (0.3) a
<i>biennial</i>	22.0 (0.4) b
Taxa	
<i>Ld</i>	30.3 (1.6) ab
<i>Ll</i> 'Honduras'	29.2 (1.6) b
<i>Ll</i> 'India'	32.0 (1.6) a
<i>Ll</i> 'K360'	30.3 (1.6) ab

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776 **Figures captions**

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779 Figure 1: Experimental design of the plot containing two parts, and a detailed description of the experimental
780 unit (replicate). The numbers **1, 2, 3** and **4** refer to the taxa (1: *Ld*, 2: *LlH*, 3: *LlI*, 4: *LlK*); and **b** are the border
781 lines, which were composed by the same *Leucaena* taxa. *Ld*: *Leucaena diversifolia*; *Ll*: *Leucaena leucocephala*
782 (*LlH*, 'Honduras'; *LlI*, 'India'; *LlK*, 'K360').

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785 Figure 2: Measurement and harvesting dates during the first three cropping years. The measurements of height
786 (*H*) and stem diameter at the base (*D*) of all the experimental units were carried out in late February to early
787 March, just before each harvest. The experimental units harvested every year after the *D* and *H* measurements
788 are indicated. Measured parameters in summer: LAI, leaf area index; SLA, specific leaf area; Ψ , predawn leaf
789 water potential.

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791 Figure 3: Total dry biomass (TDW \pm SE) accumulated throughout the first three years of the study for *L.*
792 *diversifolia* and *L. leucocephala*. (Left) Accumulated biomass for the four taxa as a whole if cutting takes place
793 only once in the indicated years (3, $R_{31}S_3$), annually (*III*, $R_{11}S_1 + R_{22}S_1 + R_{33}S_1$), and assuming annual cutting
794 the 3rd year ($R_{32}S_1$) after a biennial cutting the first time ($R_{21}S_2$) (*2I*, $R_{2,1}S_2 + R_{32}S_1$). Different letters indicate
795 significant differences between the cutting treatments ($p = 0.031$). (Right) Accumulated biomass of the four
796 *Leucaena* taxa for the *III* cutting treatment. Different letters indicate significant differences between taxa ($p <$
797 0.001) at the end of the third year after planting. *Ld*: *Leucaena diversifolia*; *Ll*: *Leucaena leucocephala* (*LlH*,
798 'Honduras'; *LlI*, 'India'; *LlK*, 'K360').

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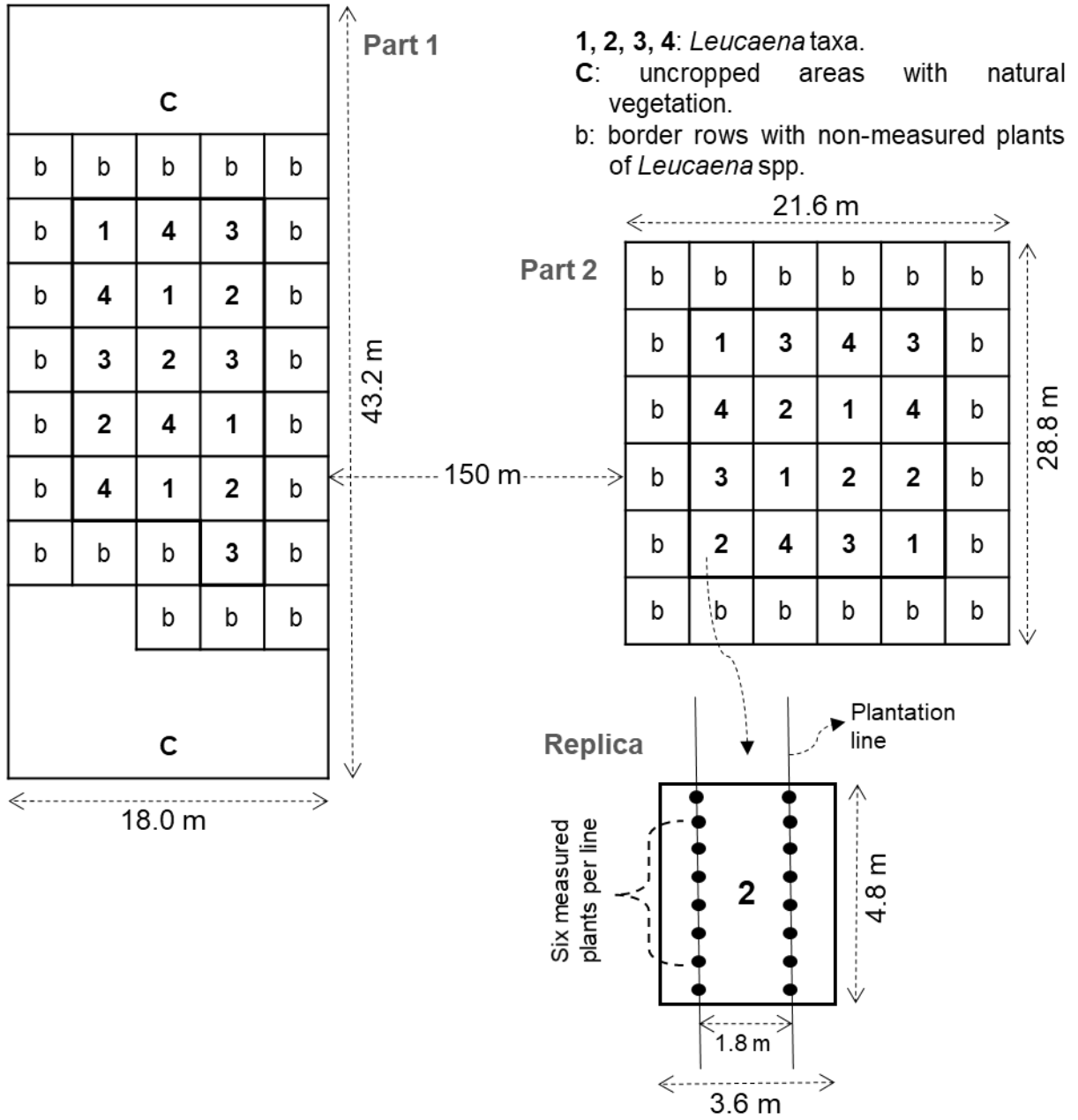
800 Figure 4: Boxplot of plant height (*H*) and stem diameter (*D*) for each taxa and cutting treatment throughout the
801 first three years of the study. Taxa: *Ld* (*L. diversifolia*), *Ll* (*L. leucocephala*) being the three taxa, 'Honduras'
802 (*LlH*), 'India' (*LlI*) and 'K360' (*LlK*). Cutting treatments: *III* [$R_{11}S_1 + R_{22}S_1 + R_{33}S_1$], *2I* [$R_{2,1}S_2 + R_{32}S_1$] and *3*
803 [$R_{31}S_3$]. Boxes are 95% and 5% percentile values, while the solid lines indicate the median. Different letters
804 depict significant differences ($p < 0.05$) between cutting treatment for each taxa.

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807 Figure 1.

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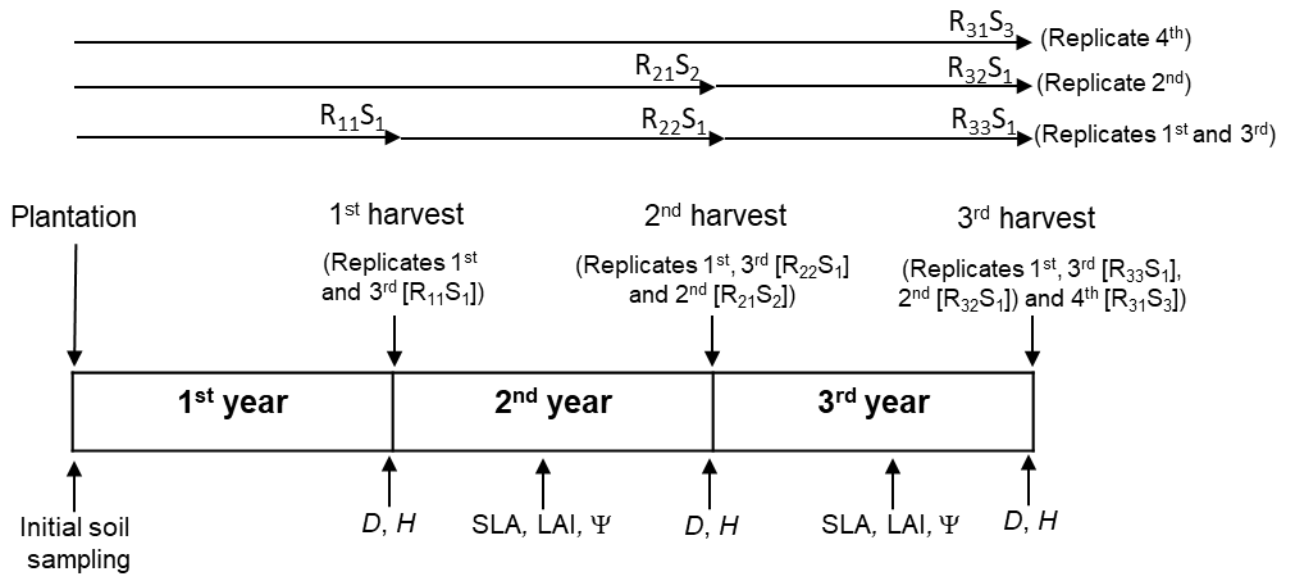


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811 Figure 2

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816 Figure 3

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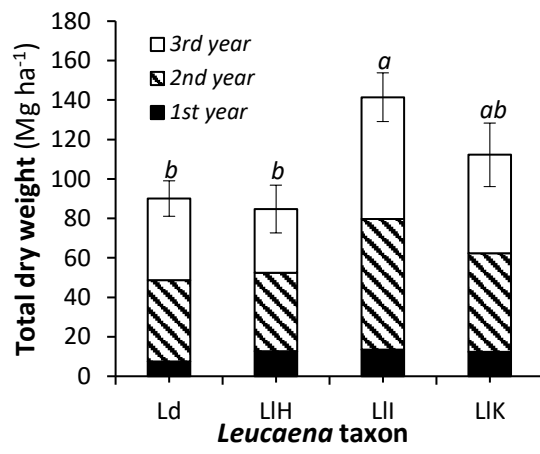
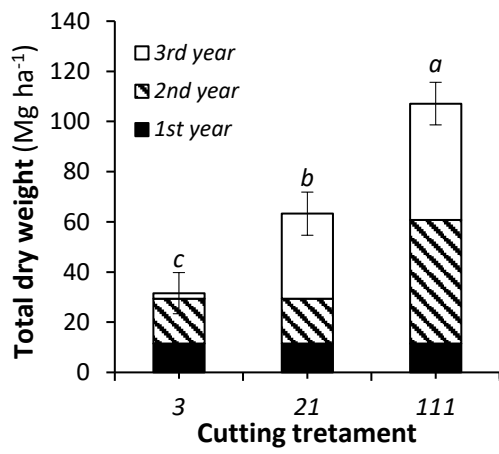
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829 Figure 4

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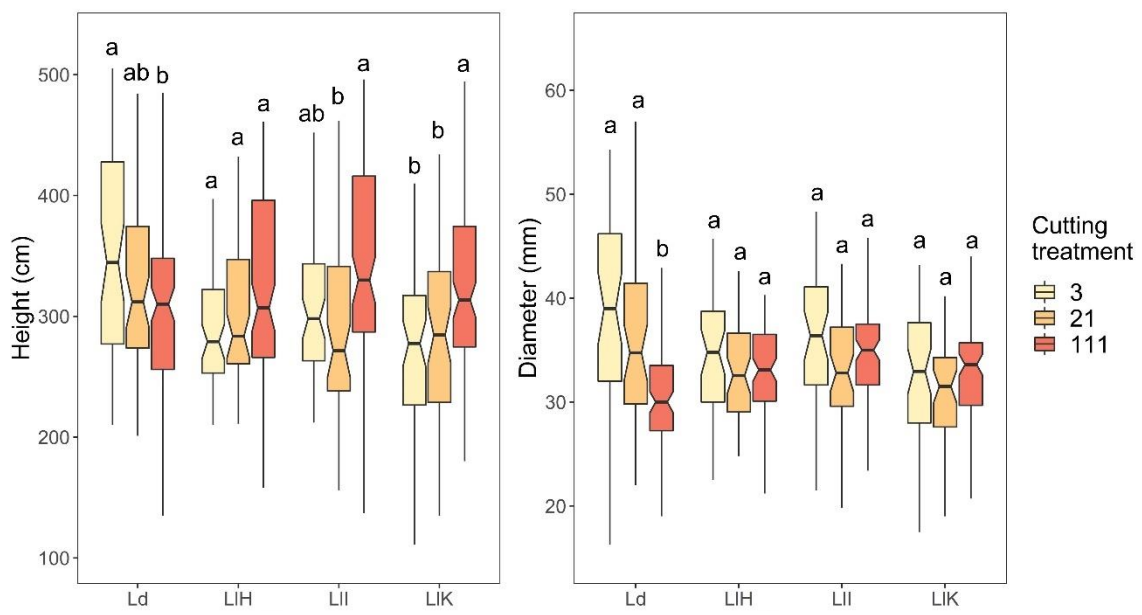
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